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## Three years of semicontinuous greenhouse gas measurements at the Puy de Dôme station (central France)

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**Abstract.** Three years of greenhouse gas measurements, obtained using a gas chromatograph (GC) system located at the Puy de Dôme station at 1465 m a.s.l. in central France, are presented. The GC system was installed in 2010 at Puy de Dôme and was designed for automatic and accurate semicontinuous measurements of atmospheric carbon dioxide, methane, nitrous oxide and sulfur hexafluoride mole fractions. We present in detail the instrumental setup and the calibration strategy, which together allow the GC to reach repeatabilities of  $0.1 \mu\text{mol mol}^{-1}$ ,  $1.2 \text{ nmol mol}^{-1}$ ,  $0.3 \text{ nmol mol}^{-1}$  and  $0.06 \text{ pmol mol}^{-1}$  for  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$ , respectively. The analysis of the 3-year atmospheric time series revealed how the planetary boundary layer height drives the mole fractions observed at a mountain site such as Puy de Dôme where air masses alternate between the planetary boundary layer and the free troposphere.

Accurate long-lived greenhouse gas measurements collocated with  $^{222}\text{Rn}$  measurements as an atmospheric tracer allowed us to determine the  $\text{CO}_2$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$  emissions in the catchment area of the station. The derived  $\text{CO}_2$  surface flux revealed a clear seasonal cycle, with net uptake by plant assimilation in the spring and net emission caused by the biosphere and burning of fossil fuel during the remainder of the year. We calculated a mean annual

$\text{CO}_2$  flux of  $1310 \pm 680 \text{ tCO}_2 \text{ km}^{-2}$ . The derived  $\text{CH}_4$  and  $\text{N}_2\text{O}$  emissions in the station catchment area were  $7.0 \pm 4.0 \text{ tCH}_4 \text{ km}^{-2} \text{ yr}^{-1}$  and  $1.8 \pm 1.0 \text{ tN}_2\text{O km}^{-2} \text{ yr}^{-1}$ , respectively. Our derived annual  $\text{CH}_4$  flux is in agreement with the national French inventory, whereas our derived  $\text{N}_2\text{O}$  flux is 5 times larger than the same inventory.

### 1 Introduction

The release of anthropogenic greenhouse gases (GHGs) into the atmosphere leads to a modification of their natural cycles and to a strong increase in atmospheric radiative forcing (Myhre et al., 2013). The Intergovernmental Panel on Climate Change (IPCC) reported that the global average temperature increased by  $0.89^\circ\text{C}$  between 1901 and 2012 (Hartmann et al., 2013) and will continue to increase during the 21st century (Collins et al., 2013). To limit the global temperature rise, most industrialized countries signed the “United Nations Framework Convention on Climate Change” (UNFCCC) treaty in 1992 to stabilize their GHG emissions between 1990 and 2000 and it entered into force in 1994. This convention was enhanced by the Kyoto Protocol, which was signed in 1997 and was ratified by 182 countries. The

countries engaged in the Kyoto Protocol aimed to reduce their national emissions of the main long-lived GHGs by 5.2 % between 2008 and 2012 compared to the emission levels of 1990. The GHGs in question are carbon dioxide ( $\text{CO}_2$ ), methane ( $\text{CH}_4$ ), nitrous oxide ( $\text{N}_2\text{O}$ ), sulfur hexafluoride ( $\text{SF}_6$ ), hydrofluorocarbons (HFCs) and perfluorocarbons (PFCs). The European Union (EU) committed itself to a reduction of its GHG emissions by 8 % for the same period. In addition, the EU aims to reduce its total GHG emissions by 20 % in 2020, relative to emissions in 1990. Despite this commitment, it is extremely difficult to validate the surface GHG fluxes on the country scale using a reliable, transparent method.

Currently, countries report their respective GHG emissions to the UNFCCC on an annual basis. These national emission inventories are based on bottom-up methods and the reliability of these national inventories strongly depends on the uncertainty attributed to each emission factor. To improve and validate the bottom-up methods, it is crucial to better characterize the biogeochemical cycles of the different GHGs, particularly as national inventories report only anthropogenic emissions to the UNFCCC. Therefore, it is important to develop new tools to quantify natural-source emissions and to provide an independent verification of the emission inventories reported to the UNFCCC.

Different methods based on atmospheric measurements can be used to estimate GHG emissions on local to regional scales. Some of these approaches couple atmospheric GHG measurements with measurements of associated atmospheric tracers of air masses, including radon-222 (Biraud et al., 2000; Schmidt et al., 2001), sulfur hexafluoride (Maiss et al., 1996) or isotopes, such as radiocarbon in  $\text{CO}_2$  (Levin and Karsten, 2007; Lopez et al., 2013). A major advantage of this “multigas” approach is that it avoids the use of complex chemistry-transport models; the tracers that are used are subject to the same atmospheric transport mechanisms as the GHGs. Nevertheless, an accurate assessment of the respective measurement station footprint is required to allocate the estimated surface fluxes to a specific region (Gloor et al., 2001).

The first atmospheric  $\text{CO}_2$  continuous measurements started in the 1950s at Mauna Loa Observatory using a nondispersive infrared (NDIR) analyzer with a repeatability better than  $0.1 \mu\text{mol mol}^{-1}$  (Keeling et al., 1976). Atmospheric  $\text{CH}_4$  monitoring began in the late 1970s using gas chromatograph (GC) systems equipped with a flame ionization detector (FID) and a nickel catalyst to enable simultaneous  $\text{CO}_2$  mole fraction detection (Rasmussen and Khalil, 1981). The repeatabilities were approximately  $10 \text{ nmol mol}^{-1}$  for  $\text{CH}_4$  measurements and  $0.7 \mu\text{mol mol}^{-1}$  for  $\text{CO}_2$  measurements. Coupling an electron capture detector (ECD) to a GC system enabled the detection of  $\text{N}_2\text{O}$  and  $\text{SF}_6$  atmospheric mole fractions with repeatabilities of approximately  $1.0 \text{ nmol mol}^{-1}$  and  $0.1 \text{ pmol mol}^{-1}$ , respectively (Weiss, 1981; Prinn et al., 1990; Maiss et al., 1996).

Subsequently, these two detectors (FID and ECD) have permitted the use of GC to analyze  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  atmospheric mole fractions simultaneously and on a semi-continuous basis. Since the 1980s, atmospheric monitoring stations that are part of the Global Atmosphere Watch (GAW) have been gradually equipped with GC systems and NDIR analyzers. The GC system technologies described above have continuously evolved to reach repeatabilities better than  $0.1 \mu\text{mol mol}^{-1}$  for  $\text{CO}_2$ ,  $2.0 \text{ nmol mol}^{-1}$  for  $\text{CH}_4$ ,  $0.3 \text{ nmol mol}^{-1}$  for  $\text{N}_2\text{O}$  and  $0.1 \text{ pmol mol}^{-1}$  for  $\text{SF}_6$ , as shown by van der Laan et al. (2009b), Thompson et al. (2009) and Popa et al. (2010).

New types of accurate instruments for  $\text{CO}_2$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$  atmospheric measurements have recently become commercially available. These instruments are based on optical technologies, including cavity ring-down spectroscopy (CRDS), Fourier transform infrared spectrometry (FTIR) and off-axis integrated cavity output spectroscopy (OA-ICOS). These recent technologies are promising for atmospheric monitoring as they offer high-frequency measurements (on the order of 1 Hz), require low maintenance and achieve equivalent or superior repeatability compared to GC systems (Yver Kwok et al., 2015; Hammer et al., 2013a). Analyzers based on the CRDS technology are generally used for  $\text{CO}_2$  and  $\text{CH}_4$  atmospheric measurements, OA-ICOS technology is used for  $\text{N}_2\text{O}$  measurements, and FTIR technology is designed to simultaneously measure  $\text{CO}_2$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$ . These new types of instruments are also more easily transportable, and Hammer et al. (2013b) have demonstrated their feasibility as “traveling” instruments. They could thus be used for comparisons and quality control purposes to ensure data compatibility across a monitoring network.

Regardless of the benefits listed above, these optical technologies cannot yet be used to measure atmospheric  $\text{SF}_6$  mole fractions;  $\text{SF}_6$  is an extremely stable GHG, having a global warming potential of 23 900 (Forster et al., 2007). In addition, most of these new technologies need to be continuously flushed, which makes it difficult to analyze flasks, in contrast to the GC systems, which employ discrete samples and can analyze four or five species simultaneously (van der Laan et al., 2009b). Studies of optical technologies are still progressing, and, as noted above, these new technologies are very promising, particularly for a dense monitoring network, such as the European infrastructure ICOS (Integrated Carbon Observation System), which is dedicated to high-precision monitoring of greenhouse gases over Europe. The CRDS technology is slowly replacing the GC systems and NDIR analyzers for  $\text{CO}_2$  and  $\text{CH}_4$  monitoring in many stations, but GC is still the reference instrument for  $\text{N}_2\text{O}$  and  $\text{SF}_6$  measurements (see WMO-GAW, 2013). Consequently, we installed a GC system in 2010 at the mountain station of Puy de Dôme (France) to monitor, with a high precision, the long-term atmospheric trend of the main four long-lived greenhouse gases.

This paper focuses on 3 years (2010 to 2013) of ambient air measurements of CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O and SF<sub>6</sub>, obtained using a GC system at the Puy de Dôme station. After a short description of the station, the detailed setup of the GC as well as the calibration strategy are addressed. A paragraph is dedicated to data quality control and atmospheric measurement comparisons, showing to what extent our measurement system are in line with the WMO recommendations (WMO-GAW, 2013). In the last part, we present and analyze our 3-year time series of ambient air measurements. Finally, we demonstrate that these time series can be used to estimate the monthly regional fluxes of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O in the catchment area of the Puy de Dôme station, using radon-222 as an atmospheric tracer.

## 2 The Puy de Dôme station

### 2.1 Site description

The Puy de Dôme station (45°46′19″ N, 2°57′57″ E) is located at the top of the Puy de Dôme volcano (1465 m a.s.l.), in Auvergne in the center of France. This station is managed by the Laboratoire de Météorologie Physique (LaMP) and is part of the Observatoire de Physique du Globe de Clermont-Ferrand (OPGC) located at Clermont-Ferrand, France. According to the French National Institute of Statistics and Economic Studies (INSEE – <http://www.insee.fr>), the ground cover of Auvergne (26 013 km<sup>2</sup>) consists mainly of meadows (36.4 %), forests (33.4 %) and arable land (17.6 %), the Puy de Dôme station being located in the center of this region. The major anthropogenic GHG sources are to the east of the station, where the town of Clermont-Ferrand is located: 10 km east of the Puy de Dôme station at an altitude of 396 m a.s.l. Clermont-Ferrand is the largest town in the region, with approximately 150 000 inhabitants.

The CITEPA (Centre Interprofessionnel Technique d'Etudes de la Pollution Atmosphérique) reports the French national GHG emissions to the UNFCCC but has also provided a regional inventory of Auvergne for the year 2007 (CITEPA, 2010). According to the CITEPA, the anthropogenic CO<sub>2</sub> emissions in Auvergne are mainly attributable to road transport and residential and industrial sectors, which represent respectively 45, 25 and 21 % of the total anthropogenic CO<sub>2</sub> emissions of the region. The agricultural sector is responsible for 90 % of total anthropogenic CH<sub>4</sub> emissions and 97 % of total anthropogenic N<sub>2</sub>O emissions in the region. Ninety percent of the SF<sub>6</sub> emissions are related to the energy transformation sector.

A military base and a telecommunication center are located 20 m northwest of the station, also on the top of the volcano. These facilities consist of a main building (20 m height) and a telecommunication antenna (89 m height). Since 2010, the only access road to the station has been closed to the public and has been replaced by a cog train.

The atmospheric research station hosts different analyzers for long-term atmospheric measurements of GHG, CO, O<sub>3</sub>, aerosol particles, radon-222, clouds microphysics and radionuclide. The station is part of the European ICOS, AC-TRIS (Aerosol particles, Clouds, and Trace gases Research InfraStructure) and EMEP (European Monitoring and Evaluation Programme) measurement networks and of the global GAW network.

### 2.2 Atmospheric conditions at the Puy de Dôme station

Meteorological parameters are monitored at the station, including wind speed, wind direction, temperature, relative humidity and atmospheric pressure. A wind shadow area between 300 and 360° is clearly observed in the wind direction due to the building and the telecommunication antenna of the military base, which both induce local turbulences. The planetary boundary layer (PBL) height, wind speed and wind direction were extracted from the European Center for Medium-range Weather Forecasts (ECMWF, 2012; Seidel et al., 2012) at a 3-hour time resolution for the years 2010 to 2012. The grid cell used for the extraction has an area of 15 km × 15 km and is centered at 45°45′ N, 3°00′ E at an altitude of 575 m a.s.l. In this study, the wind direction from ECMWF was used as the reference because the wind direction provided by the meteorological sensor is influenced by the local turbulences caused by the telecommunication antenna located at the military base. The average difference in wind speed between the meteorological station and the ECMWF data was  $3.4 \pm 4.3 \text{ m s}^{-1}$ , the wind speed measured by the sensor was higher because the sensor is located at a higher elevation than the grid cell used for the ECMWF extraction (1465 m a.s.l. compared to 575 m a.s.l.). Therefore, the wind speed from the meteorological sensor was used to correct the ECMWF data. The PBL height and the wind direction from ECMWF were interpolated using a linear regression fit to obtain a 1-hour time resolution.

The Puy de Dôme station is primarily influenced by winds from a southwest direction (48.2 % of the time) with a mean wind speed of  $8.4 \text{ m s}^{-1}$ . The wind blows from the Clermont-Ferrand sector (45–135°) only 7.7 % of the time, with a mean speed of  $4.2 \text{ m s}^{-1}$ . The PBL height analysis revealed that the Puy de Dôme station is in the free troposphere during more than 70 % of the time and up to 81 % of the time during winter.

Back trajectories were calculated using the Lagrangian dispersion model Flexpart version 8.2.3, based on ECMWF ERA-Interim data at a horizontal resolution of  $1^\circ \times 1^\circ$ , with 60 vertical levels and a 3-hour temporal resolution. Eight particles were released every 15 min (96 particles every 3 h) in a 3-D box centered around the Puy de Dôme station (from lower left corner 45.76° N, 2.95° E to upper right corner 45.78° N, 2.97° E; between 1400 and 1500 m a.s.l.) with a lifetime of 3 days. This simulation was performed for particles arriving at the station between 1 January 2010 and 31

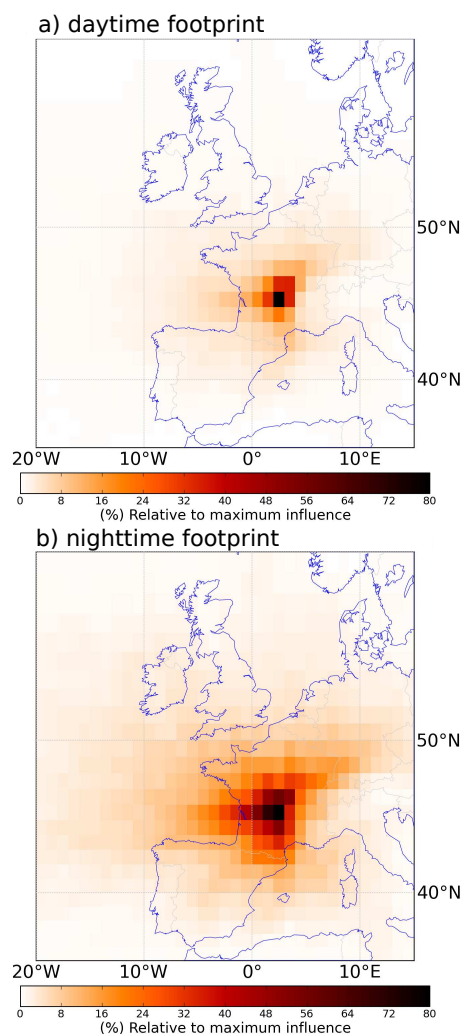
December 2012. The footprints were computed on a  $1^\circ \times 1^\circ$  horizontal grid, following the method described by Lin et al. (2003) and taking into account the planetary boundary layer height at each particle location. We considered that a particle is influenced by surface emissions from one grid cell when its elevation is under the PBL height and that its influence is inversely proportional to the PBL height. The maps presented in Fig. 1 show the footprints for air masses arriving at the station between 14:00 and 16:00 UTC, when the PBL is usually well developed (Fig. 1a), and between 22:00 and 06:00 UTC, when the PBL is below 1400 m and the station is within the free troposphere (Fig. 1b). The grid cell influence is represented as a relative influence compared to the maximum value (in percent). On both maps, the station is located within the black grid cell. By analyzing back trajectories from HYSPLIT over 437 days at Puy de Dôme (i.e., from Atlantic and continental western Europe areas), Boulon et al. (2011) showed that 87 % of air masses reaching the station are from the west. A statistical analysis of back trajectories over 4 years conducted by Venzac et al. (2009) demonstrated that winter air masses reaching Puy de Dôme travel over longer distances from the west than summer air masses.

### 3 Instrumental setup

The GHG observations at Puy de Dôme started in 2000 with continuous  $\text{CO}_2$  measurements using a nondispersive infrared (NDIR) spectrometer. Since 2001, a pair of flasks has been sampled once a week by the LaMP team and analyzed by GC for  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  mole fractions and by a mass spectrometer for  $\delta^{13}\text{C}$  and  $\delta^{18}\text{O}$  in  $\text{CO}_2$  at the LSCE in Gif-sur-Yvette, France. In 2010, the GC system for semicontinuous measurements of  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  was installed at the station. In 2011, the NDIR spectrometer was replaced by a CRDS for continuous  $\text{CO}_2$  and  $\text{CH}_4$  measurements. Since 2002, the station has also been equipped with a radon-222 ( $^{222}\text{Rn}$ ) analyzer based on the active deposit method. These instruments are housed in a regulated-temperature room, the inlet lines being located on the roof of the station, 10 m a.g.l. This section focuses on the setup of the GC system running at the Puy de Dôme station since July 2010.

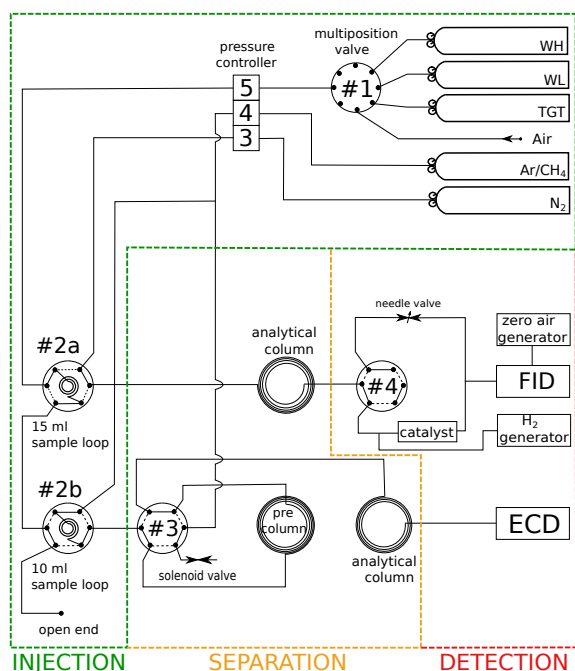
#### 3.1 Description of the GC system

The GC system installed at the Puy de Dôme station is a commercial HP-6890N from Agilent that was modified and optimized at the LSCE for automatic and semicontinuous atmospheric measurements of  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  mole fractions in dry ambient air (Lopez, 2012). Similar instrument configurations are installed at the Gif-sur-Yvette and Trainou stations in northern France (Lopez et al., 2012; Schmidt et al., 2014).



**Figure 1.** Footprint of the Puy de Dôme station from the Lagrangian dispersion model Flexpart (a) during daytime (14:00 to 16:00 UTC), when the PBL is usually well developed, and (b) during nighttime (22:00 to 06:00 UTC), when the station is usually in the free troposphere.

The ambient air is pumped from the roof of the station (pump KNF: PMF 1433-811) through a 10 m long Dekabon tube with an outside diameter of 1/2 in. ( $\sim 1.27$  cm). Three filters (140, 40 and  $7\ \mu\text{m}$  TF series from Swagelok) are placed in series to protect the pump and the analysis system from dust and aerosol particles. After passing the pump, the ambient air is pressurized and dried in two steps. First, the air passes through a commercial decanting bowl (40 mL volume) placed in a refrigerator set at  $5^\circ\text{C}$  for preliminary drying. The water accumulated in the decanting bowl is flushed out every 6 hours for 10 s by opening a solenoid valve. In a second step, the ambient air passes through a glass trap in an ethanol bath that is maintained at  $-55^\circ\text{C}$  by a cryocooler (Thermo Neslab CC-65). The dew point of the air going in the instrument is approximately  $-50^\circ\text{C}$ . The glass trap is



**Figure 2.** Schematic of the GC system setup (gas flow) at the Puy de Dôme station.

changed during the weekly maintenance of the station. An electronic box is used to regulate the refrigerator temperature and to open and close the solenoid valve of the decanting bowl. This box also records the temperatures of the fridge, the ethanol bath and the room. In case of power failure, the entire GC system is connected to an uninterruptible power supply (UPS) that allows the system to run for a few hours.

The GC system consists of an injection part, a separation part and a detection part. These three different parts are indicated by different colors in Fig. 2. For analysis, an air sample is first filled into the two sample loops. The sample is then pushed by different carrier gases to the chromatographic columns, where the species are separated. Finally, CO<sub>2</sub> (via a nickel catalyst) and CH<sub>4</sub> are detected by a FID and a microelectron capture detector ( $\mu$ ECD) is used to detect N<sub>2</sub>O and SF<sub>6</sub>. One injection and analysis requires 5.4 min.

### 3.1.1 Sample analysis

The injection part (outlined by green line in Fig. 2) consists of an eight-port microelectronic valve no. 1 (model DC8WE from Valco vici, Switzerland) that enables the selection of the samples to be analyzed (ambient air or gas cylinders). The selected sample is injected into the system via an electronic pressure control (EPC-Aux5) through two sample loops located in the room. The sample loops are placed in series on two six-port two-way Valco valves (no. 2a and no. 2b). The sample loop for CO<sub>2</sub> and CH<sub>4</sub> analysis has a volume of 15 mL (sample loop on valve no. 2a), and the one for N<sub>2</sub>O and SF<sub>6</sub> analysis has a volume of 10 mL (sample loop on

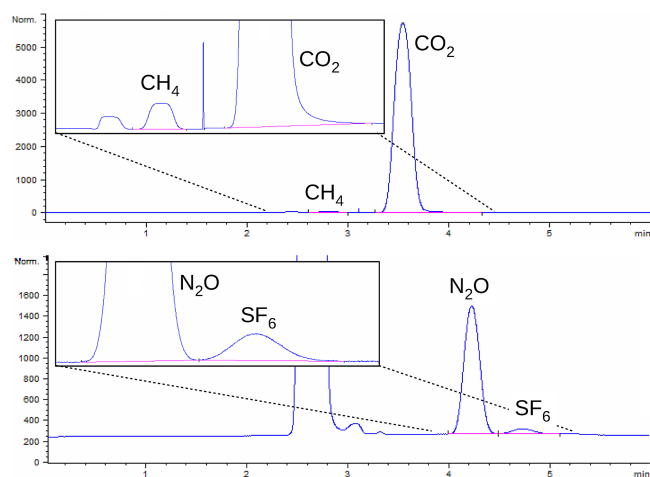
valve no. 2b). They are both flushed with the sample gas for 0.75 min at a flow rate of 180 mL min<sup>-1</sup> (corresponding to a pressure of 2.5 psi on Aux5). Before the injection, the two sample loops are equilibrated at temperature and atmospheric pressure for 0.5 min by setting Aux5 to 0 psi. After equilibration, the samples are injected into the columns with the carrier gases by opening valves no. 2a and no. 2b. The carrier gas used for the FID is N<sub>2</sub> (purity > 99.9999 %), whereas a mixture of argon and methane (95/5 %, ECD quality) is used for the  $\mu$ ECD. A purifying cartridge (Aeronex) is placed after each carrier gas cylinder.

The columns used to separate the different molecules are placed in an oven maintained at 80 °C (see the section outlined by the yellow line in Fig. 2). A Hayesep-Q (12' × 3/16" SS, mesh 80/100) analytical column is used for CO<sub>2</sub> and CH<sub>4</sub> separation. For N<sub>2</sub>O and SF<sub>6</sub> separation, a pre-column Hayesep-Q (4' × 3/16" SS, mesh 80/100) and an analytical column Hayesep-Q (6' × 3/16" SS, mesh 80/100) are used. The pre-column is back-flushed between 0 and 0.75 min and between 3.7 and 5.4 min with a 100 mL min<sup>-1</sup> flow rate of the carrier gas to eliminate heavy electrophilic molecules from the system to avoid an eventual pollution of the analytical column, which might induce an increase in the  $\mu$ ECD baseline. Between 0.75 and 3.7 min, the N<sub>2</sub>O and SF<sub>6</sub> molecules are injected first into the pre-column and then into the analytical column, where separation occurs. The analytical column is directly connected to the  $\mu$ ECD. The N<sub>2</sub>O and SF<sub>6</sub> retention times are 4.3 and 4.8 min, respectively.

The CH<sub>4</sub> and CO<sub>2</sub> molecules are detected by an FID and a Ni-catalyst used to reduce CO<sub>2</sub> to CH<sub>4</sub>. Methane molecules elute after 2.7 min and are injected directly into the detector for analysis. Once the CH<sub>4</sub> molecules are released from the analytical column, the Valco valve no. 4 is opened to connect the nickel catalyst, allowing CO<sub>2</sub> molecules to be reduced to CH<sub>4</sub> to enable CO<sub>2</sub> detection by the FID. The retention time of CO<sub>2</sub> is 3.5 min. The FID temperature is controlled at 300 °C, and the flame is fed with hydrogen at a flow rate of a 65 mL min<sup>-1</sup> (provided by an NM-H2 hydrogen generator from F-DBS) and zero air at a flow rate of a 400 mL min<sup>-1</sup> (provided by a combination of a compressor from June-Air and a 75–82 air zero generator from Parker-Balston). Hydrogen is also used for CO<sub>2</sub> reduction over the Ni catalyst. The typical efficiency of the catalyst is 97 % in the CO<sub>2</sub> atmospheric mole fraction range.

Figure 3 shows the typical chromatograms obtain by the FID and  $\mu$ ECD detectors. The top panel presents the FID's response in picoampere. The spike observed between the CH<sub>4</sub> peak and the CO<sub>2</sub> peak in the close-up panel is caused by the opening of valve no. 4. The bottom panel of Fig. 3 presents the  $\mu$ ECD's response in Hz. The first large peak observed at approximately 2.7 min is the O<sub>2</sub> peak, which is followed by the N<sub>2</sub>O peak and finally by the SF<sub>6</sub> peak.





**Figure 3.** Typical chromatograms obtained by the FID (top panel) and by the  $\mu$ ECD (bottom panel).

### 3.1.2 Analysis management

Data acquisition, valve opening and closing and temperature regulation of the GC system are entirely processed by Chemstation software (version A.10.02, Agilent). This software allows controlling the GC system parameters through the so-called “methods”. A typical method is configured to do the following:

- control the temperature of the detectors, the catalyst and the oven;
- regulate the flows of the sample (via Aux5), the carrier gases (via Aux3 and Aux4),  $H_2$  and zero air, all via five distinct EPCs;
- schedule the opening and closing of the four six-port two-way Valco valves, controlled via the internal events output GC connector;
- choose the position of the eight-port microelectronic Valco valve, controlled via the external events output GC connector; and
- integrate the results of the analysis (via the chromatograms).

A typical method is presented in Table 1 and corresponds to one analysis of a chosen sample. Table 2 summarizes the GC system setup used between 2010 and 2013. A sequence lasting 3 days is designed by the sequential arrangement of methods which enables the automatic selection of ambient air and calibration gas measurement. The created sequence runs in a loop mode.

The FID and  $\mu$ ECD signals (see Fig. 3) are expressed in picoampere and hertz, respectively. The peak integrations (area and height) of the different chromatograms are automatically computed by the Chemstation software at the end

of each method and the integration results are stored in “.txt” files. The repeatability of our GC system (see Sect. 3.3) is improved when the peak areas for  $CO_2$ ,  $CH_4$  and  $N_2O$  and the peak heights for  $SF_6$  are used. Once a day, the integration results are transferred to and stored in the LSCE database via ftp, and the mole fractions of the analyzed samples are automatically calculated. Three to five times each week, a trained operator evaluates the performances of the GC through a dedicated graphical application, enabling graphics of the instrument parameters to be drawn (see Sect. 3.3). Based on these graphics, flags are manually assigned to the data.

### 3.1.3 Calibration strategy

The GC system is calibrated using a two-point calibration strategy. Two working standards containing a known amount of  $CO_2$ ,  $CH_4$ ,  $N_2O$  and  $SF_6$  in synthetic air (matrix of  $N_2$ ,  $O_2$  and Ar) are used. The mole fractions of the trace gases in the two working standards are selected to bracket the typical ambient air mole fractions observed at the Puy de Dôme station and are referred to as working standard high (WH) and working standard low (WL). These gas mixtures are used to fill 40 L aluminum cylinders (Luxfer) to 200 bar by Deuste Steininger (Mühlhausen, Germany). All working standards are calibrated at LSCE against the laboratory standard scale of the World Meteorological Organization (WMO scale) provided by the Central Calibration Laboratories (CCL) of the National Oceanic and Atmospheric Administration (NOAA). The calibration scales currently used are WMO-X2007, NOAA-04, NOAA-2006A and NOAA-2006 for  $CO_2$ ,  $CH_4$ ,  $N_2O$  and  $SF_6$ , respectively (Zhao and Tans, 2006; Hall et al., 2007; Dlugokencky et al., 2005).

The response function of  $\mu$ ECD for  $N_2O$  analysis is non-linear, especially in the range below and above the tropospheric values (see Schmidt et al., 2001; van der Laan et al., 2009b; Lopez et al., 2012). The nonlinearity of our  $\mu$ ECD was tested by analyzing five cylinders calibrated by the CCL on the NOAA-2006A scale and with  $N_2O$  mole fractions between 302.00 and 338.04 nmol mol<sup>-1</sup>. In this small mole fraction range, which is important for our measurements, a two-point calibration describes the response function of our  $\mu$ ECD sufficiently. It compares very well with an exponential fit through five cylinders, with an average difference of  $0.01 \pm 0.13$  nmol mol<sup>-1</sup>. This result confirms that a two-point calibration strategy is well adapted to correct for the  $\mu$ ECD nonlinearity in atmospheric mole fraction ranges. Similar tests demonstrated that the FID response is linear in the atmospheric mole fraction range for  $CO_2$  and  $CH_4$  measurements.

The two working standards (WH and WL) are analyzed every 30 min to correct for atmospheric (temperature and pressure) changes as well as instrumental drifts, enabling the analysis of five samples between each calibration. The lifetime of our standards is approximately 3 years using this calibration strategy. To limit the risk of drift, each working standard must be replaced before reaching 30 bar pressure. At

**Table 1.** Measurement method used for the GC system at the Puy de Dôme station. The “On” position on the valves presented here corresponds to the dashed lines in Fig. 2 and the “Off” position correspond to the solid lines.

Time (min)	Parameter	Value	Comments
0.00	Aux 3	45.0 psi	Carrier gas pressure for N <sub>2</sub>
0.00	Aux 4	20.0 psi	Carrier gas pressure for Ar/CH <sub>4</sub>
0.00	Aux 5	2.5 psi	Sample pressure
0.00	Valve no. 2	On	Flush of the sample loops
0.00	Valve no. 3	On	Backflush of the pre-column (N <sub>2</sub> O/SF <sub>6</sub> )
0.75	Aux 5	0.0 psi	Sample pressure
0.75	Valve no. 3	Off	Stop of pre-column backflushing
1.25	Valve no. 2	Off	Sample injection
3.10	Valve no. 4	On	Injection of CO <sub>2</sub> into the catalyst
3.60	Aux 3	0.0 psi	Carrier gas pressure for N <sub>2</sub>
3.70	Valve no. 3	On	Backflush of the pre-column
5.30	Valve no. 4	Off	Opening of the catalyst valve
5.40	Aux 3	45.0 psi	Carrier gas pressure for Ar/CH <sub>4</sub>

**Table 2.** GC system equipment and temperature and flow rate settings.

Detector	FID (CO <sub>2</sub> /CH <sub>4</sub> )	$\mu$ ECD (N <sub>2</sub> O/SF <sub>6</sub> )
Carrier gas	N <sub>2</sub> cylinder (purity > 99.9999 %) + purifier	Ar/CH <sub>4</sub> cylinder (ECD quality) + purifier
Flow rate	100 mL min <sup>-1</sup>	45/65 mL min <sup>-1</sup>
Loop sample volume	15 mL	10 mL
Oven temperature	80 °C	80 °C
Pre-column		Hayesep-Q 4' × 3/16" SS, 80/100
Analytical column	Hayesep-Q 12' × 3/16" SS, 80/100	Hayesep-Q 6' × 3/16" SS, 80/100
Detector temperature	300 °C	395 °C
Catalyst temperature	390 °C	
Gas supply	H <sub>2</sub> generator: 60 mL min <sup>-1</sup> Zero air generator: 400 mL min <sup>-1</sup>	

the end of their use at the station, all working standards are recalibrated at LSCE to verify their stability over their lifetimes. At the Puy de Dôme station, the first set of working standards was replaced on 25 April 2013. Reanalysis of the standards at LSCE revealed mean differences (2010–2013) of  $-0.11 \mu\text{mol mol}^{-1}$ ,  $-0.03 \text{ nmol mol}^{-1}$ ,  $-0.1 \text{ nmol mol}^{-1}$  and  $0.0 \text{ pmol mol}^{-1}$  for CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O and SF<sub>6</sub>, respectively. The observed differences are not statistically significant except in the case of CO<sub>2</sub>. The calibration cylinders drifted by 0.08 and  $0.15 \mu\text{mol mol}^{-1}$  over their lifetime for CO<sub>2</sub>. The CO<sub>2</sub> data presented in this paper are not corrected for the observed drift on the order of  $0.03 \mu\text{mol mol}^{-1} \text{ yr}^{-1}$ . The first measurement period (July 2010 to 24 April 2013) is called “period A” in this paper, and the second measurement period (after the change of the working standards) is called “period B” (from 25 April 2013 to 30 June 2013). The mole fractions of the working standards used at Puy de Dôme are presented in Table 3.

**Table 3.** Trace gas mole fractions of GC working standards used during period A (July 2010 to April 2013) and period B (May and June 2013).

Species	Period A – 33 months		Period B – 2 months	
	WH	WL	WH	WL
CO <sub>2</sub> ( $\mu\text{mol mol}^{-1}$ )	425.10	372.45	449.60	363.31
CH <sub>4</sub> ( $\text{nmol mol}^{-1}$ )	2179.90	1732.99	2083.38	1663.52
N <sub>2</sub> O ( $\text{nmol mol}^{-1}$ )	340.90	322.93	348.03	326.51
SF <sub>6</sub> ( $\text{pmol mol}^{-1}$ )	10.05	5.38	9.86	5.86

### 3.2 Other instrumentation

#### 3.2.1 Flask sampling unit

A flask sampling unit was installed at the Puy de Dôme station in 2002 for weekly sampling. It consists of a pump that pressurizes two 1 L glass flasks placed in series to 1 bar relative pressure. They are flushed for 15 min with dry ambient



air prior the pressurization. Ambient air is dried in a distinct cooling trap maintained in the same ethanol bath as the GC trap (see Sect. 3.1). The flasks are then shipped and analyzed at LSCE by a GC system for CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O and SF<sub>6</sub> and by a Finnigan MAT-252 isotope mass spectrometer for CO<sub>2</sub> isotopic composition (<sup>13</sup>C and <sup>18</sup>O). The calibrations for trace gas analysis are performed in the same manner as presented in Sect. 3.1.3, and the results are stored in the same database.

### 3.2.2 Radon-222 measurement system

Radon-222 (<sup>222</sup>Rn) is a radioactive noble gas ( $T_{1/2}$  = 3.8 days) and is part of the radioactive decay chain of uranium-238. Uranium-238 in the earth's crust results in the emission of <sup>222</sup>Rn by the earth's surface. Atmospheric radon-222 activity has been monitored at the Puy de Dôme station since 2002. The analyzer is based on the active deposit method, which consists of alpha decay counting of <sup>222</sup>Rn's solid short-lived daughters: <sup>218</sup>Po, <sup>214</sup>Pb and <sup>214</sup>Bi. The measurement technique has been described in detail by Polian et al. (1986) and Biraud et al. (2000). To avoid the loss of the solid <sup>222</sup>Rn daughters, the inlet line is a 6 m long straight metal tube with a 31 mm outside diameter. During the first years of measurements, the inlet line was frequently contaminated by room air, and only measurements after October 2006 can be used. Schmidt (1999) estimated a radioactive disequilibrium (see Turner, 1964) at the Schauinsland station (Germany; 47°54' N, 7°54' E; 1205 m a.s.l.) of  $1.15 \pm 0.14$ . This value was independently confirmed by Xia et al. (2010). The Puy de Dôme station and the Schauinsland station are two medium-elevation mountain stations, having the same geographical environment. They are both frequently above the continental boundary layer, especially in winter. Based on the similarities between the Puy de Dôme station and the Schauinsland station, the measured (<sup>222</sup>Rn) activity at Puy de Dôme has been corrected for the radioactive disequilibrium by using the same value ( $1.15 \pm 0.14$ ).

### 3.2.3 CO<sub>2</sub> continuous measurements by in situ NDIR

Continuous CO<sub>2</sub> measurements with a NDIR gas analyzer at Puy de Dôme began in 2000. The instrument is an integrated system constructed around a LI-COR NDIR (Li-6252 NDIR, LI-COR Inc., Nebraska, USA) analyzer optical bench. The CO<sub>2</sub> measurements are based on the difference in absorption of infrared radiation passing through two cells: the reference cell and the sample cell. Infrared radiation is transmitted through both cell paths, and the analyzer signal is proportional to the difference in absorption between both cells. The measurement frequency is 1 Hz, and the cell flow is typically 20 mL min<sup>-1</sup> for the sampling cell and 15 mL min<sup>-1</sup> for the reference cell, which is continuously flushed with a reference gas. The calibration strategy is based on four cylinders calibrated on the WMO-X2007 scale. The calibration is performed twice a year by analyzing each calibration cylinder

30 times for 10 min. Data are then corrected using a quadratic fit.

Ambient air is pumped from the roof to the instrument through a 3/8 in. (~ 0.95 cm) outside diameter Dekabon line. The air is dried by passing through a glass trap maintained in a cold ethanol bath (see Sect. 3.1). Ambient air is analyzed for 50 min following the analysis of the reference cylinder for 10 min, the latter passing through both cells at the same time. The NDIR spectrometer was replaced by a CRDS analyzer in April 2011.

### 3.2.4 CO<sub>2</sub> and CH<sub>4</sub> continuous measurements by in situ CRDS

The CRDS analyzer (Picarro G1301) was installed in April 2011. It continuously and simultaneously measures CO<sub>2</sub>, CH<sub>4</sub> and H<sub>2</sub>O atmospheric mole fractions. We use four calibration cylinders spanning the atmospheric range of 366 to 453 μmol mol<sup>-1</sup> for CO<sub>2</sub> and 1722 to 2107 nmol mol<sup>-1</sup> for CH<sub>4</sub>. The cylinders are calibrated at the LSCE laboratory on the WMO-X2007 and NOAA-04 scales. The instrument calibration is performed automatically every 15 days by injection, according to the following scheme: 4 times for each calibration cylinder for 30 min, beginning with the one with the lowest mole fractions and ending with the one with the highest mole fractions. For each calibration cylinder, the entire first injection and the first 15 min of the subsequent injections are automatically rejected because of equilibration time. The instrument calibration takes 8 h and a linear fit is applied to compute the analyzer response. A target gas is automatically injected into the CRDS every 10 h for 30 min and, again, the first 15 min are automatically rejected. The SDs at 1 sigma of the target gas analysis over 1 year are 0.02 μmol mol<sup>-1</sup> for CO<sub>2</sub> and 0.14 nmol mol<sup>-1</sup> for CH<sub>4</sub>.

Ambient air is injected into the CRDS from the roof (through a 3/8 in. Dekabon line equipped with three filters of 140, 40 and 7 μm) using a pump located after the optical cavity. To avoid the risk of bias caused by the interference of water vapor and trace gases in the CRDS (Chen et al., 2010), the ambient air is dried prior to its injection into the CRDS by the drying system presented in Sect. 3.1. A problem occurred in the cavity in August 2011, and the CRDS measurements were stopped until April 2012. The instrument has been returned to the manufacturer for repair.

### 3.3 Quality control of the GC system and comparisons with different analyzers

A target gas (TGT) is injected into the GC system once an hour for quality control. The target gas cylinder is a 40 L cylinder filled with dry ambient air at Gif-sur-Yvette. After a stabilization time of at least 1 month, the cylinder is analyzed at LSCE using the laboratory primary standards, and CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O and SF<sub>6</sub> mole fraction values are assigned to the cylinder (see Table 4).

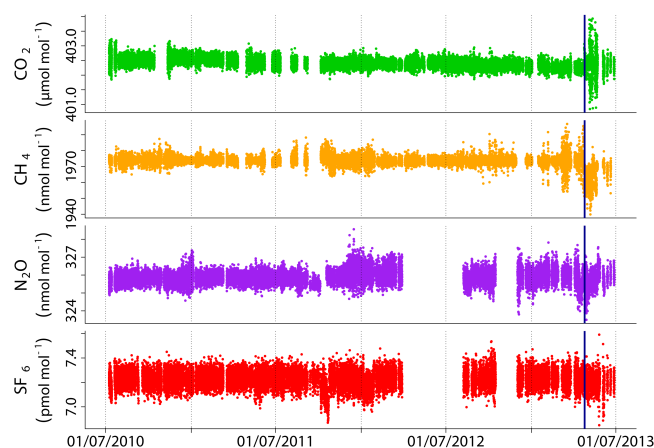
**Table 4.** Assigned target gas values with the respective mole fractions measured using the GC system at the Puy de Dôme station over period A and period B. The assigned values were measured by the GC system at the LSCE central lab against WMO calibration gases.

Species	Assigned values	Period A	Period B
CO <sub>2</sub> (μmol mol <sup>-1</sup> )	402.57 ± 0.07	402.42 ± 0.15	402.38 ± 0.46
CH <sub>4</sub> (nmol mol <sup>-1</sup> )	1973.87 ± 0.73	1973.81 ± 2.12	1963.72 ± 6.64
N <sub>2</sub> O (nmol mol <sup>-1</sup> )	325.71 ± 0.23	325.90 ± 0.35	325.76 ± 0.47
SF <sub>6</sub> (pmol mol <sup>-1</sup> )	7.23 ± 0.04	7.23 ± 0.06	7.20 ± 0.07

Figure 4 shows the time series of the target gas analysis from July 2010 to June 2013. The target gas cylinder was not changed over the 3 years. The different data gaps observed in CO<sub>2</sub> and CH<sub>4</sub> between March and October 2011 were caused by problems with the hydrogen generator (leaks in the electrolysis cell). The large data gaps in N<sub>2</sub>O and SF<sub>6</sub> between April and August 2012 were due to a problem with the power supply of Valco valve no. 2b. The reproducibility (computed here as the SD at 1 sigma of the target analysis over 1 year of measurement) and the typical short-term repeatability (SD of the target analysis over 24 h) are presented in Table 5.

The vertical dark blue lines in Fig. 4 indicate when the working standards were changed and separate the two measurement periods: period A and period B (see Sect. 3.1.3). The average mole fractions of the TGT measurement during period A and period B are presented in Table 4 together with the respective assigned mole fractions. The measured CH<sub>4</sub>, N<sub>2</sub>O and SF<sub>6</sub> mole fractions agreed well with the assigned mole fractions during period A (considering the uncertainties). The difference between the CO<sub>2</sub> assigned and measured values in period A and period B was 0.15 and 0.19 μmol mol<sup>-1</sup>, respectively, confirming the consistency between the two scales used in periods A and B. This agreement when using two different calibration scales shows that the problem is probably due to an error in the value attributed to this target gas. During period B, the average CH<sub>4</sub> mole fraction of the TGT gas was lower by approximately 10 nmol mol<sup>-1</sup> compared to period A. This decrease was most probably due to a micro-leak in the WH line that affects only the CH<sub>4</sub> mole fractions. Target data as well as ambient air measurements for this period should be recalibrated by applying a one-point calibration to the four trace gases. This micro-leak was detected and fixed in September 2013, when the target cylinder was replaced by a new one.

In addition to internal quality control performed via the target gas analysis, comparisons of in situ ambient air analysis, flask analysis and cylinder analysis performed by different analyzers are relevant. These comparisons enable the validation of the scale consistency between different instruments and the detection of possible leaks or biases introduced by the inlet lines. The WMO-GAW gives recommendations on the scientific level of compatibility for such a comparison



**Figure 4.** Time series of the target gas measured with the GC system at Puy de Dôme. The vertical blue lines in each panel correspond to the date of the working standards change.

**Table 5.** Reproducibility and typical short-term (24 h) repeatability of the GC system at Puy de Dôme, both within 1 sigma.

Species	Reproducibility	Short-term repeatability
CO <sub>2</sub> (μmol mol <sup>-1</sup> )	0.14	0.1
CH <sub>4</sub> (nmol mol <sup>-1</sup> )	2.12	1.2
N <sub>2</sub> O (nmol mol <sup>-1</sup> )	0.34	0.3
SF <sub>6</sub> (pmol mol <sup>-1</sup> )	0.06	0.06

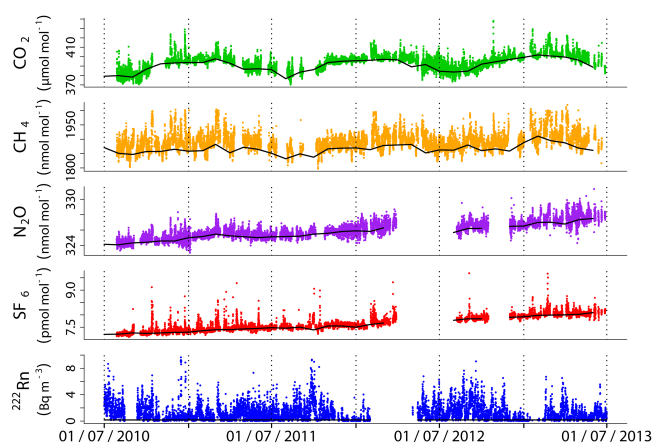
in the Northern Hemisphere (WMO-GAW, 2013). These levels are ±0.1 μmol mol<sup>-1</sup> for CO<sub>2</sub>, ±2.0 nmol mol<sup>-1</sup> for CH<sub>4</sub>, ±0.1 nmol mol<sup>-1</sup> for N<sub>2</sub>O and ±0.02 pmol mol<sup>-1</sup> for SF<sub>6</sub>.

The mean differences between the in situ GC system measurements and weekly flask sampling, in situ NDIR measurements and in situ CRDS measurements are summarized in Table 6. These differences are calculated from the hourly mean measurements. A 2-sigma filter was applied to the differences to flag the eventual outliers. Comparisons between GC and NDIR were based on 9 months of overlapping measurements (July 2010 to April 2011) and revealed a mean CO<sub>2</sub> difference (GC minus NDIR) of  $-0.14 \pm 1.78$  μmol mol<sup>-1</sup>. Comparisons between GC and CRDS were based on 20 months of overlapping measurements, from April 2011 to July 2013 with a break between August 2011 and April 2012. The average differences (GC minus CRDS) were  $0.21 \pm 0.78$  μmol mol<sup>-1</sup> for CO<sub>2</sub> and  $-0.64 \pm 5.46$  nmol mol<sup>-1</sup> for CH<sub>4</sub> over the total overlapping measurements. Because the CRDS instrument was stopped for several months and shipped to the manufacturer in 2011, we compared the results before and after its repair. In the first overlapping measurement period (April to August 2011), the differences were  $-0.13 \pm 0.61$  μmol mol<sup>-1</sup> and  $-1.27 \pm 3.49$  nmol mol<sup>-1</sup> for CO<sub>2</sub> and CH<sub>4</sub>, respectively.

In the second overlapping measurement period (April 2012 to July 2013) the  $\text{CH}_4$  difference decreased to  $-0.26 \pm 5.02 \text{ nmol mol}^{-1}$ , whereas the  $\text{CO}_2$  difference increased to  $0.28 \pm 0.75 \text{ } \mu\text{mol mol}^{-1}$ . Over the second comparison period, the observed  $\text{CO}_2$  difference remained constant with time and did not depend on atmospheric mole fractions. The inlet lines, including the pumps and the dryer systems, were tested for 3 weeks: a common inlet line for ambient air measurements has been used for the GC and CRDS (the GC one, described in Sect. 3.1). During these 3 weeks of testing, the difference between GC and CRDS remained constant and equal to  $0.28 \text{ } \mu\text{mol mol}^{-1}$ , confirming that the inlet lines did not cause the observed bias. Even after changing the GC working standards in late April 2013, the  $\text{CO}_2$  difference was still observed. This observed difference is stable over time and is not concentration dependent. A second experiment was to analyze the second set of working standards on the CRDS. The results showed a difference between the assigned value (at LSCE) and the CRDS of  $0.03 \text{ } \mu\text{mol mol}^{-1}$  on the WL and of  $0.34 \text{ } \mu\text{mol mol}^{-1}$  on the WH.

Comparisons between the GC system and the flask analysis or comparison cylinders provide information on the scale consistency between different laboratories. For the four analyzed long-lived GHGs, the comparisons between GC in situ measurements and flask analyses reached the desirable comparison levels (see Table 6 for more details). The two GC measurements bracketing each sampled flask measurements are linearly interpolated in order to match the time of the flask sampling. The Puy de Dôme station also participates in the “Cucumbers comparison programme” (<http://cucumbers.uea.ac.uk/>) in the framework of the European Union CarboEurope project (2000–2005), EU IMECC (2007–2011) and InGOS (2011–2015) infrastructure projects. Three cylinders are alternately analyzed on the GC at Puy de Dôme and at the Gif-sur-Yvette, Trainou (France), Kasprowy Wierch (Poland) and Hegyhatsal (Hungary) stations. Table 6 presents the mean differences between the average analysis of the three comparison cylinders at Puy de Dôme and Gif-sur-Yvette (LSCE) between 2011 and 2013.

The different comparison methods presented in this section showed that the GC system installed at the Puy de Dôme station matches the WMO-GAW recommendations for  $\text{CH}_4$  and  $\text{SF}_6$  measurements. The  $\text{CO}_2$  comparison shows different results depending on the method used. The recommendations are met by considering the comparison with cylinders or flasks, whereas they are not if we consider only the comparison between in situ instruments. The WMO-GAW recommendations concerning the  $\text{N}_2\text{O}$  measurements are ambitious considering the repeatability obtained with our GC. The  $\text{N}_2\text{O}$  measurements are not in line with the WMO-GAW recommendation but we note that the different comparison methods used show differences lower than our instrumental repeatability.



**Figure 5.** Hourly mole fractions of  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  atmospheric ambient air and hourly activities of  $^{222}\text{Rn}$  at Puy de Dôme from July 2010 to June 2013. The black lines are the respective monthly GHG background mole fractions of the Puy de Dôme measurements (between 22:00 and 06:00 UTC) and the  $^{222}\text{Rn}$  background activity at Mace Head (Ireland).

## 4 Results and discussions

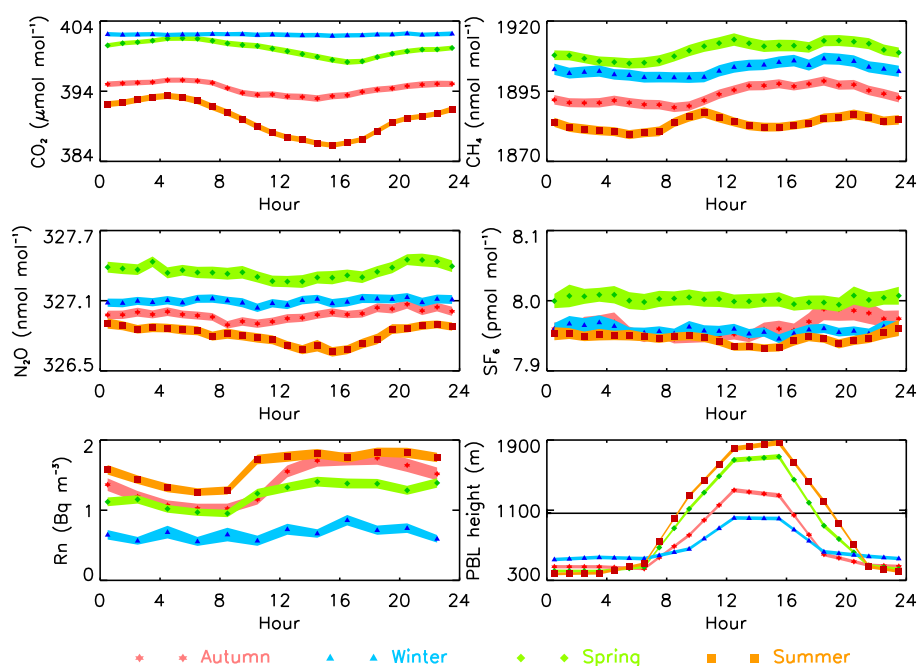
### 4.1 Three years of ambient air measurements

Figure 5 shows the hourly time series of  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  ambient air mole fractions together with the  $^{222}\text{Rn}$  activities at Puy de Dôme from July 2010 to the end of June 2013. The different gaps observed in the atmospheric GHG time series are explained in Sect. 3.3. These time series are presented with the respective monthly background values (black lines). The monthly background values were calculated from the monthly average nighttime mole fractions (between 22:00 and 06:00 UTC), when the station is above the PBL (see Fig. 6). The hourly  $^{222}\text{Rn}$  activities are presented in the last panel of Fig. 5 and varied between 0 and  $9 \text{ Bq m}^{-3}$  over the 3 years of measurements. From February 2012 to the end of April 2012, the computer for  $^{222}\text{Rn}$  data acquisition had hardware and software problems, resulting in the observed gap.

Figure 6 presents the mean diurnal cycles per season of  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$ ,  $\text{SF}_6$ ,  $^{222}\text{Rn}$  and for the PBL height (from ECMWF) from June 2010 to June 2013. The GHG diurnal cycles were computed from the detrended hourly time series to the reference of 1 January 2013. The mean yearly increase rate was subtracted from the time series, before computing the seasonal mean diurnal cycles. To represent the thickness of the PBL relative to the ground level, we used the altitude of Clermont-Ferrand (396 m a.s.l.) as the reference altitude to plot the PBL height, Clermont-Ferrand being located at the lowest altitude of the ECMWF extracted grid cell. The horizontal solid black line on the PBL height panel in Fig. 6 gives the altitude of the station above Clermont-Ferrand and enables a quick observation of whether the sta-

**Table 6.** Results of the comparisons between in situ measurements obtained using the GC system, NDIR, CRDS and flask measurements. The mean differences in the cylinder analysis at the Puy de Dôme and Gif-sur-Yvette stations are presented in the last column. Flasks and cylinders were analyzed at the LSCE central lab at Gif-sur-Yvette.

Date Comparisons	July 2010–July 2013 GC in situ – flasks Average No. of flasks		July 2010–April 2011 In situ: GC–NDIR	April 2011–July 2013 In situ: GC–CRDS	July 2010–July 2013 Cylinders: PUY–GIF
CO <sub>2</sub> (μmol mol <sup>-1</sup> )	0.11 ± 1.19	55	-0.14 ± 1.78	0.21 ± 0.78	-0.02 ± 0.11
CH <sub>4</sub> (nmol mol <sup>-1</sup> )	0.04 ± 4.30	57	n/a	-0.64 ± 5.46	0.64 ± 0.26
N <sub>2</sub> O (nmol mol <sup>-1</sup> )	0.12 ± 0.55	47	n/a	n/a	0.21 ± 0.47
SF <sub>6</sub> (pmol mol <sup>-1</sup> )	-0.01 ± 0.07	53	n/a	n/a	0.03 ± 0.03



**Figure 6.** Mean diurnal cycles of CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, SF<sub>6</sub>, and <sup>222</sup>Rn together with the planetary boundary layer height (relative to Clermont-Ferrand altitude – 396 m a.s.l.) at the Puy de Dôme station for each season. Trace gas mole fractions were detrended based on 1 January 2013.

tion is within or above the PBL. The mean diurnal cycles of the PBL height exhibited the same pattern for each season, with an increase in height from 06:00 to 12:00 UTC followed by a stable height until 15:00 UTC. After 15:00 UTC, the PBL height began to decrease, reaching a minimum after 21:00 UTC. On a mean annual scale, the GC sampled the trace gases within the PBL between 09:00 and 18:00 UTC with an enlarged time step in summer and a narrower time step in winter. In winter, the Puy de Dôme station is often above the PBL during several consecutive days.

The mean diurnal cycles of the long-lived GHGs and <sup>222</sup>Rn observed in this study are typical of mountain sites, as previously described by Schmidt et al. (1996) for the Schauinsland station (1205 m a.s.l.) or Necki et al. (2003) for the Kasprowy Wierch station (Poland, 1987 m a.s.l.). The PBL height is a key atmospheric factor, particularly for mountain sites where measurements alternate between the

free troposphere and the PBL. The atmospheric mole fraction variabilities of trace gases are generally larger in the PBL because of the combination between its diurnal variability and the emissions from surface sources. Due to its short radioactive lifetime, <sup>222</sup>Rn cannot accumulate in the free troposphere, in contrast to other long-lived trace gases. The mean diurnal cycles of <sup>222</sup>Rn (Fig. 6) exhibited larger variations in summer, when the PBL height is maximal and the inlet line of the station alternates between the PBL during daytime and above during nighttime. In winter, <sup>222</sup>Rn activities are lower than in summer because the station is usually in the free troposphere.

The mean diurnal cycles for CO<sub>2</sub> and CH<sub>4</sub> exhibit different shapes. As for the <sup>222</sup>Rn activities, the CH<sub>4</sub> mole fractions at the Puy de Dôme station are higher in the afternoon, when the PBL is well developed, compared with the nighttime mole fractions. On a yearly average, CH<sub>4</sub> afternoon mole fractions

are  $3.3 \text{ nmol mol}^{-1}$  higher than the nighttime mole fractions. We observed an opposite diurnal cycle trend for  $\text{CO}_2$ : the biosphere is a sink for  $\text{CO}_2$  during the daytime and counterbalances the atmospheric effects. This is seen clearly during summertime, when the photosynthetic activity is maximal: the amplitude of the  $\text{CO}_2$  diurnal cycle is  $7.2 \mu\text{mol mol}^{-1}$ , with a minimum mole fraction of around 16:00 UTC. The  $\text{CO}_2$  mole fractions are maximal in winter when the biosphere acts as a  $\text{CO}_2$  source, mainly driven by soil respiration.

The amplitudes of the  $\text{N}_2\text{O}$  and  $\text{SF}_6$  diurnal cycles are very small and nearly undetectable, except for the  $\text{N}_2\text{O}$  in summer which exhibits an amplitude of  $0.25 \text{ nmol mol}^{-1}$ . The  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  mole fractions are largest in spring and lowest during summertime because their respective mole fractions are mainly driven by the PBL height and the associated vertical mixing.

## 4.2 The marine boundary layer reference

In this section, the background mole fractions of recorded trace gases at the station (see Sect. 4.1) are compared with the respective marine boundary layer references (MBLRs). Here, the MBLRs are the monthly zonal average trace gas mole fractions for  $45.5^\circ \text{ N}$  computed from NOAA measurements (Dlugokencky et al., 2013a, b). They were retrieved from the Global Monitoring Division of the NOAA Earth System Research Laboratory. Figure 7 shows the differences between the monthly mean background mole fractions at Puy de Dôme (nighttime values between 22:00 and 06:00 UTC) and the respective monthly MBLRs for  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$ . These comparisons enable the direct quantification of the influence of sources and sinks on trace gases at the station relative to oceanic air masses. These differences are called continental offsets. The  $\text{CO}_2$  continental offset has negative values in spring, indicating the influence of the continental biosphere, which acts as a sink. During summer, autumn and winter, the offsets are positive, revealing the importance of continental fossil fuel and biospheric sources in the Puy de Dôme catchment area. The continental offsets are usually positive for  $\text{CH}_4$  and always positive for  $\text{N}_2\text{O}$ , indicating the strong influence of agricultural sources (see Sect. 2.1) in the Puy de Dôme footprint. Finally, the  $\text{SF}_6$  offset varied between  $-0.10$  and  $+0.12 \text{ pmol mol}^{-1}$ , which represents the same order of magnitude as the GC measurement repeatability. In addition, Fig. 7 shows the monthly  $^{222}\text{Rn}$  continental offset (for nighttime selection data between 22:00 and 06:00 UTC) at the Puy de Dôme station relative to marine air. The marine air  $^{222}\text{Rn}$  reference was computed from the mean activity of 15 years of measurements during maritime background conditions at the European background site of Mace Head (see Bousquet et al., 1996) and is equal to  $168 \text{ mBq m}^{-3}$ .

## 4.3 The radon tracer method

### 4.3.1 Method

Once emitted by soils,  $^{222}\text{Rn}$  is an excellent tracer of continental air masses due to its physical and chemical properties. Thus the radon tracer method (RTM) has been used in numerous atmospheric studies to estimate trace gas surface emissions on local to regional scales. Detailed descriptions of this method are given in the following studies: Schmidt et al. (2001); Hammer and Levin (2009); Yver et al. (2009); van der Laan et al. (2009a).

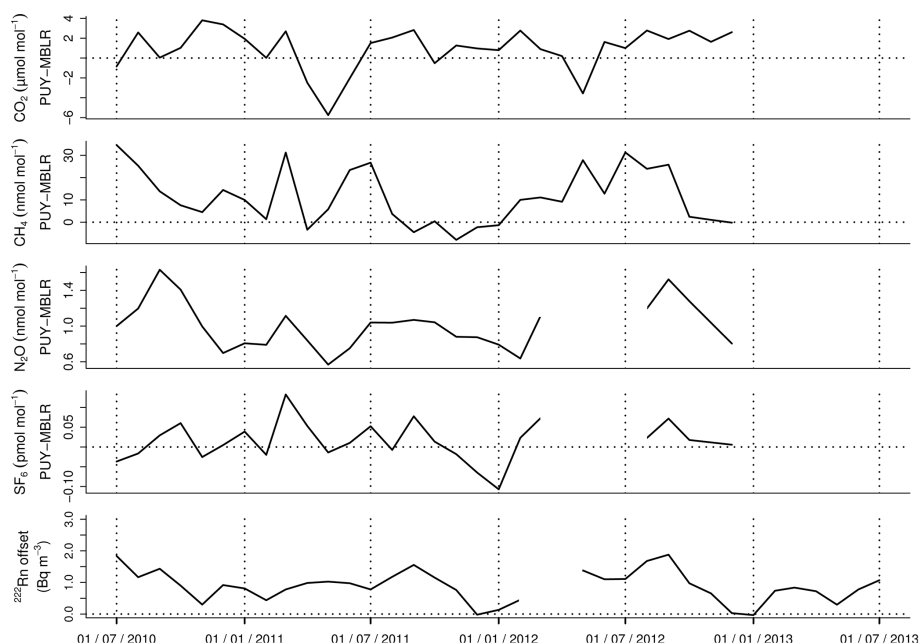
The RTM is based on Eq. (1), where  $J_x$  and  $J_{\text{Rn}}$  are the respective fluxes of a trace gas  $x$  and  $^{222}\text{Rn}$ . The  $\Delta C_x$  and  $\Delta C_{\text{Rn}}$  terms are the temporal variations in the trace gas  $x$  mole fraction and in the  $^{222}\text{Rn}$  activity over a period  $\Delta t$ . Finally,  $\lambda_{\text{Rn}}$  is the  $^{222}\text{Rn}$  decay constant.

$$J_x = J_{\text{Rn}} \frac{\Delta C_x}{\Delta C_{\text{Rn}}} \left( 1 - \frac{\lambda_{\text{Rn}} C_{\text{Rn}}}{\frac{\Delta C_{\text{Rn}}}{\Delta t}} \right) \quad (1)$$

As shown in Fig. 6, the diurnal variations in trace gases at Puy de Dôme are very weak, which makes difficult to correctly assess the  $\Delta C_x$  and  $\Delta C_{\text{Rn}}$  terms on a daily basis. In this study, we apply the RTM approach presented by Schmidt et al. (2003), in which the  $\text{CO}_2$  fluxes at the Schauinsland station were calculated using the monthly  $\text{CO}_2$  and  $^{222}\text{Rn}$  continental offsets (relative to the MBLR). As presented in Sect. 4.2, the continental offsets of trace gases reflect the source and/or sink influence at a continental site relative to a maritime background. In this study, the terms  $\Delta C_x$  and  $\Delta C_{\text{Rn}}$  (see Eq. 1) were calculated as the monthly offsets of trace gases and radon-222, respectively.

The term in brackets in Eq. (1) corresponds to the radioactive decay correction factor which depends on the mean residence time of air masses over the European continent before reaching the station. In Sect. 2.2, it was shown that most of the air masses arriving at the station are from the western part (oceanic air masses) and have an average wind speed of  $8.4 \text{ m s}^{-1}$ . Based on this, it takes at least 10 h for the oceanic air masses to reach the station from the closest oceanic coast. Considering also that some of the air masses are from other directions, we estimate an average transit time for the air masses arriving at the station of 1 day, leading to a decay correction of 0.91.

The  $^{222}\text{Rn}$  emission rate from continental surfaces strongly depends on the type and on the nature of the soils. A study of Karstens et al. (2015) provides a monthly  $^{222}\text{Rn}$  emission map at a resolution  $0.083^\circ \times 0.083^\circ$  over Europe. The assessment of this map takes into account the soil types and properties, the  $^{238}\text{U}$  soil content and the soil moisture evolution over time. According to the mean nighttime footprint at the Puy de Dôme station (see Fig. 1b), we extracted the monthly  $^{222}\text{Rn}$  average emission from this map for a  $300 \text{ km} \times 300 \text{ km}$  region centered on the Puy de Dôme



**Figure 7.** Differences between the monthly background at Puy de Dôme and the respective monthly MBLR (Dlugokencky et al., 2013a, b) at 45.5° N latitude for CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O and SF<sub>6</sub>. The last panel is the <sup>222</sup>Rn offset relative to marine air.

station (U. Karstens and I. Levin, personal communication, 2014). Over the years 2010 to 2012, the <sup>222</sup>Rn fluxes range between 75 and 172 Bq m<sup>-2</sup> h<sup>-1</sup>, with minimums in winter, when the soil is wet or frozen.

#### 4.3.2 Uncertainties

The uncertainties of the radon tracer method presented above result from errors in the <sup>222</sup>Rn exhalation rate, errors in the  $\Delta C_x$  and  $\Delta C_{Rn}$  terms and error in the decay correction term (see Eq. 1). This section describes how these errors have been assessed to derive a mean relative uncertainty of the flux estimation of each trace gas.

A systematic assessment of the <sup>222</sup>Rn exhalation rate is quite difficult. The mean ratio of the spatial variability within the extracted area (300 km × 300 km region centered on the Puy de Dôme station) to the mean flux is 30 % (Karstens et al., 2015, and U. Karstens and I. Levin, personal communication, 2014). This number is used as a first approximation of the <sup>222</sup>Rn exhalation rates uncertainties. This estimate does not include systematic errors and therefore is likely an underestimate. Uncertainties in the  $\Delta C_x$  term have been assessed from the MBLRs and from the background mole fraction uncertainties. The monthly MBLR uncertainties for CO<sub>2</sub> and CH<sub>4</sub> were provided by NOAA; mean uncertainties over the measurement period have been taken into account and are equal to 0.6 μmol mol<sup>-1</sup> and 5.1 nmol mol<sup>-1</sup>, respectively. We used a mean N<sub>2</sub>O MBLR uncertainty of 0.3 nmol mol<sup>-1</sup>, which was estimated and provided by E. J. Dlugokencky (personal communication, 2014). The uncertainties regard-

ing the background mole fractions at Puy de Dôme were derived directly from the respective GC repeatabilities (see Table 5). These last two error sources were combined to give the mean absolute continental offset ( $\Delta C_x$ ) over the entire measurement period. Thus, the mean relative uncertainties in the continental offsets are estimated to be 31, 39 and 42 % for  $\Delta C_{CO_2}$ ,  $\Delta C_{CH_4}$  and  $\Delta C_{N_2O}$ , respectively. The <sup>222</sup>Rn instrument has an absolute error of ±20 % for continental measurements (Biraud et al., 2000). Based on the same approach as for the  $\Delta C_x$  term, a constant uncertainty of 28 % has been attributed to the <sup>222</sup>Rn continental offset term. Finally, Schmidt et al. (2003) reported an error of 7 % in the decay correction (term in brackets in Eq. 1) estimated at Schauinsland.

These uncertainties were combined using the square root over the quadratic sum. The mean relative flux uncertainties derived for our RTM approach were 52, 57 and 59 % for CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O, respectively. The uncertainties estimated here using this continental RTM approach are larger than those found by Biraud et al. (2000), Schmidt et al. (2001), van der Laan et al. (2009a) and Lopez et al. (2012), which are all close to 35 % for the CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O flux estimates. The uncertainties presented here are mainly driven by the continental offset uncertainties.

The uncertainties in the SF<sub>6</sub> emissions are up to 300 %. Therefore, we do not present any SF<sub>6</sub> emissions in this study.



### 4.3.3 Estimation of GHG surface fluxes in the Puy de Dôme catchment area

Continental  $\text{CO}_2$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$  surface fluxes at the Puy de Dôme station were calculated using the radon tracer method. As shown by Gloor et al. (2001), knowledge of the station footprint is an important parameter in interpreting the large time variability in a trace gas mole fraction observed at a measurement station. Figure 1b shows the integrated nighttime footprint (22:00 to 06:00 UTC) of the Puy de Dôme station between 2010 and 2013 (see Sect. 2.2 for more details). The station is mainly influenced by regional air masses, which are well distributed all around the station during nighttime, when the measurements are usually performed in the free troposphere.

The calculated monthly fluxes are presented in Fig. 8 together with the hourly  $^{222}\text{Rn}$  exhalation rate (U. Karstens and I. Levin, personal communication, 2014) at Puy de Dôme. The units used to express the trace gas fluxes are  $\text{t km}^{-2} \text{ month}^{-1}$  for  $\text{CO}_2$  and  $\text{CH}_4$  and  $\text{kg km}^{-2} \text{ month}^{-1}$  for  $\text{N}_2\text{O}$ . Because no  $^{222}\text{Rn}$  activities were recorded between January and the end of April 2012, no fluxes could be derived from the RTM. The vertical grey lines on each curve are the absolute uncertainties calculated in Sect. 4.3.2.

The  $\text{CO}_2$  fluxes integrate the signals from all  $\text{CO}_2$  sources and sinks in the nighttime footprint of the station. These are the contributions of the biosphere (emissions and uptakes) and of the anthropogenic emissions (fossil fuel and biofuel). The derived  $\text{CO}_2$  fluxes present negative values in spring, emphasizing the net uptake by the plant assimilation with a monthly average value between April and June of  $-435 \pm 226 \text{ t CO}_2 \text{ km}^{-2} \text{ month}^{-1}$  in the station catchment area. Schmidt et al. (2003) calculated  $\text{CO}_2$  fluxes at the Schauinsland station from 1980 to 2000 and observed a long-term monthly mean  $\text{CO}_2$  uptake between May and June of  $147 \text{ t CO}_2 \text{ km}^{-2}$ , with a maximum uptake of  $550 \text{ t CO}_2 \text{ km}^{-2}$  in the spring of 1989. These values are of the same order of magnitude as the estimations of this study. In summer, fall and winter, the fluxes are positive, indicating that the  $\text{CO}_2$  signal is dominated by the biospheric (predominantly soil respiration) and fossil fuel emissions. The monthly average  $\text{CO}_2$  flux over the total measurement period in the Puy de Dôme station footprint is  $109 \pm 57 \text{ t CO}_2 \text{ km}^{-2} \text{ month}^{-1}$ . The CITEPA (French emission inventory) provides only anthropogenic emissions for Auvergne; these were  $21 \text{ t CO}_2 \text{ km}^{-2} \text{ month}^{-1}$ . Our approach cannot separate biospheric sources and fossil fuel sources; therefore, a direct comparison between the atmospheric approach and the emission inventory is not possible.

The  $\text{CH}_4$  fluxes exhibit large variabilities, with monthly values between  $-1.04 \pm 0.59$  and  $1.65 \pm 0.93 \text{ t CH}_4 \text{ km}^{-2} \text{ month}^{-1}$ . Negative values occurred in April, September and November 2011 due to biases in the calculated background induced by the many data gaps during the months considered. Therefore,

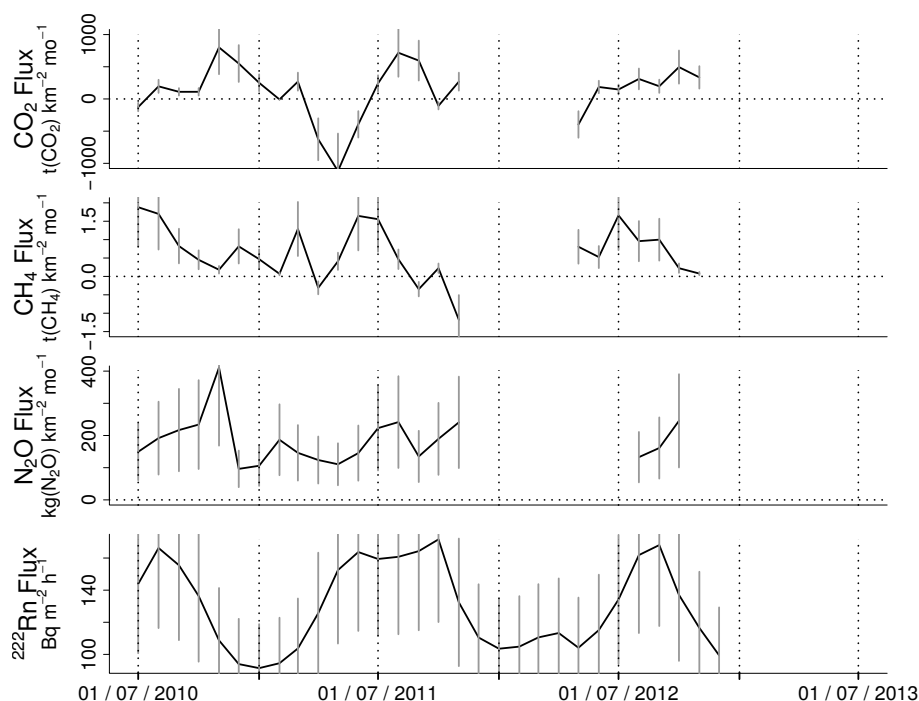
these negative fluxes are not taken into account in the average flux calculation. The average  $\text{CH}_4$  emission was  $7.0 \pm 4 \text{ t CH}_4 \text{ km}^{-2} \text{ yr}^{-1}$  over the total measurement period. The  $\text{N}_2\text{O}$  estimate emissions varies between  $84 \pm 50$  and  $360 \pm 213 \text{ kg N}_2\text{O km}^{-2} \text{ month}^{-1}$ , with a mean annual emission of  $1760 \pm 1040 \text{ kg N}_2\text{O km}^{-2} \text{ yr}^{-1}$ .

Several studies have used the radon tracer method to estimate  $\text{CH}_4$  and/or  $\text{N}_2\text{O}$  emissions over western Europe (Biraud et al., 2000; Schmidt et al., 2001; Lopez et al., 2012). The results of these estimations are summarized in Table 7 together with the  $\text{CH}_4$  and  $\text{N}_2\text{O}$  emissions estimated by this study and the estimate provided by the CITEPA for Auvergne. The estimates of  $\text{CH}_4$  emissions in the cited literature agree well over western Europe, with the exception of the estimation of van der Laan et al. (2009a), who calculated much higher  $\text{CH}_4$  emissions for the Netherlands. Following Fig. 1, the grid cells contributing the most to the signal measured at the station at night cover an area of approximately  $300 \text{ km} \times 300 \text{ km}$ . Auvergne covers an area of approximately  $150 \text{ km} \times 250 \text{ km}$  also centered on the station, but the neighboring regions are also rural areas presenting roughly the same land cover and similar GHG fluxes, which allows a direct comparison between the fluxes estimated by our atmospheric approach and those estimated by the CITEPA for Auvergne.

The CITEPA estimates a yearly  $\text{CH}_4$  emission of  $6.0 \text{ t CH}_4 \text{ km}^{-2} \text{ yr}^{-1}$  for Auvergne, indicating good agreement between the inventory and the atmospheric approach. However, our study overestimates the  $\text{N}_2\text{O}$  emissions by a factor of 5 compared with the CITEPA estimations. The  $\text{N}_2\text{O}$  fluxes are mainly driven by agricultural sources in Auvergne (CITEPA, 2010), and such fluxes strongly depend on the soil characteristics, soil temperature, and amount and type of fertilizer used. Thus, soil  $\text{N}_2\text{O}$  emissions are extremely heterogeneous, which explains the distribution of results obtained in the different studies cited in Table 7. The high  $\text{N}_2\text{O}$  emissions observed in this study may be attributable to the influence of a local agricultural source. These differences are also linked to significant uncertainties, which are strongly driven by the small continental offsets between 0.6 and  $1.5 \text{ nmol mol}^{-1}$ . Despite this difference, the atmospheric approach presented provides an independent estimation of GHG emission over the station footprint as well as new information on flux seasonality.

## 5 Conclusions

Semicontinuous measurements of four long-lived GHGs at Puy de Dôme started in 2010 with the installation of a GC system. This GC is designed to automatically measure  $\text{CO}_2$ ,  $\text{CH}_4$ ,  $\text{N}_2\text{O}$  and  $\text{SF}_6$  atmospheric mole fractions. We described in details three methods which have been used for comparison purposes. They are based on a direct comparison between two in situ analyzers, flask measurements and cylin-



**Figure 8.** Monthly  $\text{CO}_2$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$  fluxes at the Puy de Dôme station derived from the radon tracer method. The last panel presents the hourly  $^{222}\text{Rn}$  exhalation rate (U. Karstens and I. Levin, personal communication, 2014). The vertical grey lines are the respective flux uncertainties.

**Table 7.** Summary of  $\text{CH}_4$  and  $\text{N}_2\text{O}$  flux estimations and their respective uncertainties over western Europe using the RTM (this study; Biraud et al., 2000; Schmidt et al., 2001; van der Laan et al., 2009a; Lopez et al., 2012; and regional emission inventory of CITEPA).

Study	Station	Catchment area	Years	$\text{CH}_4$	$\text{N}_2\text{O}$
				$\text{t CH}_4 \text{ km}^{-2} \text{ yr}^{-1}$	$\text{kg N}_2\text{O km}^{-2} \text{ yr}^{-1}$
This study	Puy de Dôme (night)	Auvergne	2010–2012	$7.0 \pm 4.0$	$1760 \pm 1040$
CITEPA (2010)	Emission inventory	Auvergne	2007	$6.0 \pm 3.0$	$320 \pm 640$
Biraud et al. (2000)	Mace Head	western Europe	1996–1997	$4.8\text{--}3.5 \pm 1.5$	$475\text{--}330 \pm 120$
Schmidt et al. (2001)	Schauinsland	western Europe	1996–1998		$1180 \pm 345$
van der Laan et al. (2009a)	Lutjewad	the Netherlands	2006–2009	$15.2 \pm 5.3$	$900 \pm 300$
Lopez et al. (2012)	Trainou (180 m a.g.l.)	central region (France)	2009–2012		$520 \pm 156$

der measurements. For  $\text{CH}_4$  and  $\text{SF}_6$ , all comparisons show that GC measurements at Puy de Dôme are in agreement with the WMO-GAW compatibility goals. For  $\text{N}_2\text{O}$ , our measurements do not match the ambitious WMO-GAW compatibility goal. For  $\text{CO}_2$ , the comparison based on ambient air flasks and reference cylinders analysis between the GCs operated at Puy de Dôme and at LSCE reaches the desirable comparison level, showing there is no bias in the scale transfer between the two sites. Nevertheless, it does not do so for the in situ comparison with other analyzers (NDIR and CRDS). The comparisons between the GC and the CRDS in situ measurements indicate a constant offset of  $0.21 \mu\text{mol mol}^{-1} \text{ CO}_2$  over 20 months of overlapping measurements. Several tests have been performed and are described in the study, but the reason for the observed constant bias is not yet clear. We are

continuing to work on this issue and are therefore aware of the order of magnitude of bias that is possible.

At stations that typically run only one analyzer, a bias of  $0.25 \mu\text{mol mol}^{-1}$  might not be detected when the target gas and the comparison cylinders yield good results. For consistency, we thus recommend using different methods based on flask or cylinder comparisons but also based on in situ comparisons to check whether the considered measurements match with the WMO-GAW recommendations.

The diurnal cycles of  $\text{CO}_2$  and  $\text{CH}_4$  observed at Puy de Dôme are mainly driven by the PLB height, and they present the typical shape of a mountain station, such as the Schauinsland or Kasprowy Wierch stations, while the  $\text{N}_2\text{O}$  and  $\text{SF}_6$  mean diurnal cycles present flat behaviors that are difficult to interpret.

Radon-222 was used in this study as an air mass tracer to estimate the monthly continental fluxes of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O relative to the maritime background layer references. We derived a yearly net emission of 1310 tCO<sub>2</sub> km<sup>-2</sup>, 7.0 tCH<sub>4</sub> km<sup>-2</sup> and 1.7 tN<sub>2</sub>O km<sup>-2</sup>. The derived CO<sub>2</sub> and CH<sub>4</sub> fluxes compare well with other European studies or with the national inventory (CITEPA). However, it remains difficult to compare the N<sub>2</sub>O fluxes with other studies due to large errors. Compared to the GC system presented in this study, the new analysis technique based on CRDS, FTIR or OA-ICOS achieve better precision and require less maintenance. Consequently, the use of these new technologies enables the development of a dense measurement network, such as ICOS, which will further improve uncertainties in the flux estimates.

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